Regional Scale Transport in a Karst Aquifer

2. Linear Systems and Time Moment Analysis

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Travel time distributions of water or tracers in conduit-type karst aquifers can be found from linear systems analysis of either tracer test data or naturally occurring fluctuations in springflow chemistry. I use the chemical fluctuations at Maramec Spring, Missouri described in paper 1 (Dreiss, this issue) and results from a previous tracer test to derive a set of kernel functions that represent regional scale transport in the karst conduit network. A single kernel is sufficient to simulate the storm-derived component of Maramec springflow, suggesting that rapid transport in the conduit network is well-approximated by temporal stationarity. Time moment analysis of the kernels leads to several conclusions. The kernel for the tracer test exhibits a larger mean residence time and much smaller variance than the kernels derived from nonpoint source recharge. Thus the tracer travel distance appears to be longer than the mean travel distance of rapid recharge and much of the variance of nonpoint source kernels apparently results from the distribution of flow path lengths to the spring. By assuming an effective transport model and comparing the moments of the empirical tracer test kernel to the moments of the impulse response of the model, I compute an effective velocity between the tracer input point and Maramec Spring of 1.3 km/day and an effective dispersivity of 0.29 km. Because the time moments of the kernels and the effective transport parameters can be computed from readily measured springflow properties, they may prove to be a convenient means for studying and comparing regional scale transport in karst aquifers.

INTRODUCTION

In many karst aquifers, sparsely distributed solution features control rates and directions of a large portion of the regional groundwater flow. Water quickly infiltrates after storm events and some of the infiltration moves rapidly through a network of solution conduits to spring outlets. In the first paper [Dreiss, this issue], I estimated amounts of recent, storm-derived water in the discharge of a spring in southeastern Missouri. Using measured chemical perturbations in the spring flow, I found that rapidly transmitted water comprised approximately 25% of the total spring discharge. Relatively dilute water began to arrive at the spring within 1 day after each large storm event. The dilute component reached a maximum during the recession of the springflow hydrograph, about 2–5 days after the events.

This type of discrete conduit flow cannot be described with continuum models for flow through porous media because the representative elementary volume for the conduit network is undefined and likely to be very large. Similarly, models of flow in individual conduits are not feasible because the location and geometry of the conduits are unknown. Here, I present an application of linear systems and time moment analysis to the spring discharge and cation concentration measurements from southeastern Missouri. The systems approach is well-suited to studies of karst aquifers because it describes regional scale transport in terms of the distribution of travel or residence times without requiring detailed knowledge of the internal structure of the aquifer.

In this paper, I treat the observed chemical fluctuations at Maramec Spring as naturally occurring, regional tracer events that represent the response of the karst aquifer to a series of nonpoint source inputs. Examples of kernel function identification are presented using five sets of data: (1) a long-term, 12-month sequence, three short-term subsets of the record, and a tracer test previously performed at the spring. I then compare time moments of the kernels and discuss them in the context of two alternative approaches for describing karst aquifers: (1) the conceptual classification scheme of diffuse and conduit-type aquifers and (2) a dual-porosity, convective-dispersive transport model.

BACKGROUND

Linear systems analysis has been used for many years in hydrology to describe rainfall-runoff relationships and in chemical engineering to interpret reactor performance [e.g., Dooge, 1973; Himmelblau and Bischoff, 1968]. Recently, several studies have proposed the use of linear transfer functions to describe regional scale solute transport in fractured aquifers and highly heterogeneous soils [Duffy and Harrison, 1987; Jury et al., 1986; Rinaldo and Gambolati, 1987]. The problems encountered in modeling these processes are similar to those in karst hydrology.

Linear systems analysis was first applied to karst aquifers by Knisel [1972] and Dreiss [1982, 1983] in studies that considered excess precipitation as an input time series and spring discharge as an output series. The identified kernel functions in these studies were analogous to instantaneous unit hydrographs for surface runoff.

Recently, Duffy and Gelhar [1986] applied spectral analysis to tritium concentration histories to identify a kernel function and phase spectra for transport to a spring in a karst region of the Swiss Alps. They then compared the frequency distribution of the empirical transfer function to that of a one-dimensional, convective-dispersive model to obtain estimates of the primary karst reservoir volume and the dispersivity of the system. The investigation used biweekly tritium measurements in a region with significant snowmelt contributions. The accuracy of the study results may have been effected by the long sampling interval [see Dreiss, this issue] and the difficulty of estimating the input series because of snowpack storage of
tritium. Nevertheless, the approach is similar to the one presented here. The major differences are that Duffy and Gelhar [1986] use frequency domain analysis of biweekly isotope data and an effective dispersion model, and this study utilizes discrete deconvolution of daily cation concentrations and time moment analysis of the resulting kernel.

**Linear Systems Analysis**

If assumptions of linearity and time invariance are made, the observed fluctuations in the chemical composition of spring flow can be analyzed by treating the karst conduit system as a linear filter which transforms a series of input stimuli into output responses (Figure 1). In this case, the input values are rapid infiltration rates into the conduit network. Outputs are flowrates of storm-derived, relatively dilute water arriving at the spring outlet.

For a linear, time-invariant system, a single empirical transfer function or kernel function, which reproduces the observed behavior of the physical system, can be identified from known or measured input and output time series. In tracer tests, where the input and output series are the mass rate of solute entering and leaving the system, respectively, the kernel function represents both the residence time distribution of the solute in the system, and the probability distribution of solute travel times.

**Kernel Function Identification**

The behavior of a linear, time-invariant system is described by the convolution integral

\[ y(t) = \int_{0}^{t} h(t - \tau) x(\tau) \, d\tau \]  

(1)

where \( y(t) \) is the continuous system response, \( x(\tau) \) is the input series, and \( h(t - \tau) \) is the kernel function [Dooge, 1973]. For discrete, noisy measurements, (1) becomes

\[ y_i = \Delta t \sum_{j=0}^{i} x_j h_{i-j} + e_i, \quad i = 0, 1, 2, \ldots, N \]  

(2)

for \( N + 1 \) output data points collected at intervals of equal length \( \Delta t \). Output values at times \( i \Delta t \) are \( y_i \). The \( x_j \) and \( h_i \) represent mean input and kernel function values over time step intervals \( \Delta t \). The error terms \( e_i \) are residuals due to measurement errors, or, if model assumptions are not strictly met, nonlinearities and temporal changes in the system.

The sensitivity of kernel function identification to inaccurate data and nonlinearities has been demonstrated by a number of researchers [e.g., Blank et al., 1971; Delleur and Rao, 1971; Laurenson and O'Donnell, 1969]. If there is error in the data, the identification problem becomes one of finding an optimum kernel function using all available information. Methods of least squares, time series analyses and mathematical programming are most commonly applied to this problem [e.g., Eagleson et al., 1966; Neuman and de Marsily, 1976; Kanasewich, 1975].

In this study, we find \( h_{i+1} \) by minimizing the sum of the square of the errors \( E \)

\[ \min E = \sum_{i=0}^{N} e_i^2 \]  

(3)

subject to the constraints that the discrete kernel is nonnegative:

\[ h_k \geq 0 \quad k = 0, 1, 2, \ldots, M \]  

(4)

where \( M \) is the memory length of the system. Equations (2) through (4) are solved using a constrained least squares procedure similar to that described by Kanasewich [1975].

When the total volumes of the input and output series are equal, the area under the kernel is unity

\[ \Delta t \sum_{k=0}^{M} h_k = 1 \]  

(5)

and the \( h_k \) represent the response of the system to an instantaneous unit input. In applications to karst aquifers, the shape of the kernel function represents the age distribution of water in the conduit network and the distribution of solute travel times for regional scale transport through the network. For example, the fraction of spring flow of age equal to or less than \( \tau \) is the integral of the kernel between 0 and \( \tau \), or approximately,

\[ \Delta t \sum_{k=0}^{\tau} h_k \]  

(6)

and the fraction older than \( \tau \) is approximately

\[ \Delta t \sum_{k=\tau+1}^{M} h_k \]  

(7)

Time moments of the kernel describe the shape of the function and the statistical properties of the travel or residence time distribution.

**Time Moment Analysis**

The method of moments is widely used in time series analyses [e.g., Priestly, 1981] and has recently been applied in studies of transport through porous media [e.g., Valocchi, 1985; Goltz, 1986]. The basic premise of time moment analysis is that the shape of a downstream response of a system observed at point \( x \) may be described by a set of moments

\[ \mu_n(x) = \int_{0}^{\infty} \tau^n \eta(x, \tau) \, d\tau \]  

(8)

where \( \mu_n(x) \) is the \( n \)th moment about the origin at location \( x \), \( \tau \) is time, and \( \eta(x, \tau) \) is the observed disturbance [Aris, 1958]. In
the context of the kernel for a karst aquifer, the $h_k$ describe the response of the conduit system to an instantaneous unit input located a distance $x$ from the spring. The zeroth order moment $\mu_0$ is the area under the kernel, which in this case is equal to 1.

The first moment of $h_k$, approximated for discrete data as

$$\mu_1 = \Delta t \sum_{k=0}^{M} k h_k$$

is the centroid of the area under the kernel. The quantity $\mu_1/\mu_0$ has dimensions of time and is the mean residence time $\bar{t}$ of water traveling to the spring.

The amount of spreading or mixing in the system response is described by the variance $\sigma^2$ of the response, which is the second central moment about $\bar{t}$

$$\sigma^2 = \Delta t \sum_{k=0}^{M} (k - \bar{t})^2 h_k$$

or nondimensionality as the coefficient of variation $C_v$, where

$$C_v = \sigma/\bar{t}$$

Values of $\sigma^2$ and $C_v$ are related to the distribution and interconnectedness of travel pathways in a system. Kernel functions for an aquifer with a narrow distribution of pathways will exhibit less variance than those for an aquifer with a broad pathway distribution. Similarly, the variance of $h_k$ will be larger for aquifers where travel pathways are highly interconnected and smaller for aquifers, such as conduit-type karst aquifers, where large quantities of water are transmitted rapidly in discrete conduits. Thus the variance should be larger for springs in less mature karstic terrains and in aquifers with large recharge areas.

The skewness of $h_k$ can be expressed as the third central moment $\Gamma$ where

$$\Gamma = \Delta t \sum_{k=0}^{M} (k - \bar{t})^3 h_k$$

or, nondimensionally, as the skewness coefficient $\gamma$

$$\gamma = \Gamma/\sigma^3$$

Skewness results from at least three phenomena: (1) increasing dispersion with time as the response passes the observation point, (2) a nonsymmetric distribution of travel distances, and (3) delay effects from storage in relatively immobile portions of the aquifer. The first of these becomes less important with increasing fluid flow velocities. The second depends on the distribution of distances between points of rapid recharge and the spring outlet. The third develops as water enters and leaves dead-end solution conduits, fine pores, and fractures in response to changes in fluid pressures in the conduits.

APPLICATIONS

Daily precipitation, spring discharge, and cation concentration measurements were made at Maramec Spring during a 375-day period between November 1985 and November 1986. Here I use those data, as well as the results from a tracer test at the spring, to identify kernel functions for transport through the karst aquifer.

1985–1986 Record

First, consider the full-length data record shown in Figure 2 in the first paper [Dreiss, this issue], in which I used spring flow and cation concentration measurements to estimate $Q_{aw}$—the mean daily flowrate of storm-derived water arriving at the spring outlet [Dreiss, this issue, Figure 6]. That series is used here as the output response $y_i$ in the kernel identification (Figure 2).

The input sequence $x_j$ is an excess precipitation series, scaled so that the total volume of the input and output series are equal. Excess precipitation in this case refers to the total measured precipitation less evapotranspiration losses computed with a Thornthwaite moisture balance [Thornthwaite and Mather, 1957]. In karst regions, this excess precipitation contributes both to surface runoff and to groundwater circulation in the conduit network. A portion of the groundwater circulation feeds to spring outlets and the remainder leaves as deep groundwater circulation out of the region. To avoid the difficulty of estimating separate surface runoff, spring flow, and deep circulation components, a constant fraction of each excess daily precipitation event is assumed to eventually reach the spring outlet. The $x_j$ are computed by scaling the daily excess precipitation series by a factor equal to the ratio of the total output volume to the total volume of excess precipitation. Thus the total volume of the scaled input series is equal to the volume of the output response and the computed kernel function represents a unit response function of the conduit network.

Initially, I identified an optimal kernel function from $x_j$ and $y_i$ using (2) through (4) with a nonnegativity constraint but no memory length specification. Figure 2 shows the computed kernel and its ability to reproduce $y_i$. The kernel was not smoothed and has an apparent memory length of 223 days and a maximum at $k = 5$ days. It simulates the output series well, with an average absolute error of 0.56 m$^3$/s, where the average error is defined as the sum of the absolute differences between $y_i$ and the simulated output values divided by $N$. The beginning of the sequence is underestimated by the kernel because the observed $y_i$ include discharge from storms that occurred prior to the start of the record.

The kernel exhibits large initial values between 0 and 40 days and a tail of smaller values between 40 and 223 days. The first moment of $h_k$, the mean residence time, is 69.1 days. The second and third moments $C_v$ and $\gamma$ are 0.72 and 0.75, respectively.

The $h_k$ for $k > 40$ days could reflect relatively long travel times for transport through the karst network. However, this tail portion of the kernel is approximately 6 months long and could also result from seasonal errors in the estimation of $x_j$ or $y_i$. For example, in computing $y_i$, prestorm Ca ion concentrations $C_{old}$ were assumed to be constant throughout the monitoring period at the maximum measured Ca concentration of 36.9 mg/L [Dreiss, this issue]. $C_{old}$ might actually change seasonally, becoming lower during times of high infiltration and higher after long periods with no infiltration. If this is the case, but only the late summer maximum $C_{old}$ has been used to compute $y_i$, the output series will be overestimated during periods of high infiltration.

To demonstrate the effect of a seasonally varying $C_{old}$ on the computed $y_i$ and $h_k$, a new series $y_i^{*}$ was formed using $C_{old}^{*}$ values that fluctuate on an annual cycle (Figure 3). $C_{old}^{*}$ was assumed to vary between the maximum measured calcium concentration, 36.9 mg/L on September 29 and 30.3 mg/L measured 182 days previously on March 31. This value was measured exactly half a year before the maximum concentration and was close to the minimum measured value in
the series. The corrected $C_{old}^*$ is a cosine function with a wavelength of 364 days and an amplitude of 3.3 mg/L:

$$C_{old}^*(i) = \bar{C} + (36.9 + \bar{C}) \cos \left[ \frac{(\text{nday} - 309)2\pi}{364} \right]$$

where $\bar{C}$ is $(36.9 + 30.2)/2$, nday is the number of days since the start of the record, and 309 is the number of days to the date of maximum concentration.

Figure 3 illustrates the new $y_i^*$ and kernel functions identified from $y_i^*$ and the original $x_i$ series. Again, the kernels were identified from (2) through (4). No memory length was specified in deriving the $h_k$ in Figure 3a. The seasonal correction in
the output series eliminates the tail in the derived kernel function except for a series of values between 175 and 230 days. These values are associated with a disproportionately large spring flow response after several relatively small precipitation events. Specifying the memory of the \( h_n \) to be 175 days or less truncates these values but does not alter the shape of the main part of the kernel (Figure 3b). In this case, the kernel has a memory of only 38 days and reproduces the output response well with the exception of the response between 175 and 230 days. The \( h_n \) derived with the memory constraint exhibit a maximum at \( k = 5 \) days, with an average absolute error of 0.50 m\(^3\)/s. The shorter memory length is reflected in the computed time moments. The \( i \) range between 11.1 and 11.8 days, \( C_r \) between 0.68 and 0.77, and \( \gamma \) between 0.83 and 1.14. This consistency suggests that a single linear kernel adequately describes transport through the system when the inputs are nonpoint source precipitation events.

**Tracer Test**

The kernels in Figure 4 represent arrival times for basin scale transport from nonpoint sources. They depict travel along numerous pathways, over a distribution of travel distances, as well as transport processes such as dispersion along individual pathways. To illustrate and compare a kernel from a point source to the nonpoint source kernels, we examined the results of a tracer test conducted in an earlier study at Maramec Spring.

In May 1982, Vandike [1984] placed Rhodamine dye in a creek bed approximately 21 km from the spring. Dye began arriving at the spring 11 and 12 days later and continued to be present in the spring water for at least 12 days. The input for the kernel identification is an instantaneous unit input at the time of the tracer injection. The output responses \( y_m \) are the mass rate of arrival of the dye, normalized by the total mass that arrived at the spring outlet,

\[
y_m = \frac{q_mC_m\Delta t}{N} \sum_{m=1}^{N} q_mC_m\Delta t
\]

where \( N \) is the total number of sampling intervals and \( m \) is the number of intervals since the start of the test. The \( q_m \) are the mean daily springflows during the \( m \)th intervals and \( C_m \) are the measured dye concentrations. In this case, \( y_m \) is a unit response to the tracer injection and

\[
h_n = y_m \quad k = m = 1, 2, \ldots, M
\]

The derived kernel and its moments are included in Figure 4 and Table 1. The mean residence time of the tracer is 16.0 days. This is 4–5 days greater than the \( i \) of the nonpoint kernels and indicates that the travel distance of the tracer was longer than the mean travel distance of rapid infiltration. \( C_r \) of the tracer kernel is 0.19, much smaller than the nonpoint source kernels, suggesting that much of the variance in the nonpoint source kernels is due to the distribution of flowpath lengths to the spring. On the other hand, the \( \gamma \) of the tracer kernel is 0.96, similar to values for the nonpoint kernels. Thus the skewness appears not to be sensitive to the travelpath distribution or the path length distribution.

**DISCUSSION**

Physical interpretations of the moments can be discussed further in the context of two alternative models for transport in karst aquifers. One approach is based on the qualitative description of karst aquifers as either diffuse or conduit types.
and an analogy between this classification scheme and a mixing cell model. The mixing cell model is not physically realistic but provides a straightforward, alternative means of describing the degree of mixing in the aquifer. A second approach is a deterministic model for transport in dual-porosity porous media similar to those used to describe solute transport in unsaturated soils. The model is more physically detailed and allows effective transport properties, i.e., effective dispersivity and velocity, to be defined and to be estimated from the time moments of the kernel functions.

Diffuse and Conduit Aquifer Classification

Shuster and White [1971] proposed a method for qualitatively classifying karst aquifers as either “diffuse” or “conduit” type aquifers depending on several properties of the aquifers and springs. In diffuse flow aquifers, springs are usually small and controlled by stratigraphic and structural features. Spring chemistry does not vary seasonally or after storm events. In conduit aquifers, the specific conductivity of springflow is highly variable, seasonally and after large storm events. Groundwater flow occurs in solution conduits and is sometimes described as “pipe flow” or “plug flow.”

This very useful conceptual method for describing karst aquifers can be interpreted in terms of the degree of mixing that occurs within the aquifer. Purely conduit-type aquifers or purely diffuse aquifers are end members of the spectrum of aquifers that actually occur in nature. In a purely diffuse flow aquifer, recharge is completely mixed with preexisting water in the aquifer, so the spring flow shows very little variation in chemistry over time. In contrast, pure conduit flow results in little mixing and therefore large temporal variation in spring flow chemistry.

Diffuse and conduit aquifers might be conceptualized as end-members of a set of mixing cell models where the degree of mixing is represented by a series of mixing cells as shown in Figure 5a. The material balance in each cell is [Himmelblau and Bischoff, 1968]

$$V' \frac{dc_i}{dt} + Qc_i = Qc_{i-1} \quad i = 1, 2, \ldots, I$$

where $c_i(t)$ is the tracer concentration in the $i$th cell, $V'$ is the volume of an individual cell, $Q$ is a constant flow rate, and $I$ is the number of cells in the model. A kernel function for a

### TABLE 1. Time Moments for Storm Sequences and Tracer Test

<table>
<thead>
<tr>
<th>Date</th>
<th>$N_i$</th>
<th>$C_{det}$</th>
<th>$M_i$</th>
<th>Average Error, m/s</th>
<th>$\bar{t}_i$</th>
<th>$C_v$</th>
<th>$\gamma$</th>
</tr>
</thead>
<tbody>
<tr>
<td>November 1985 to</td>
<td>375</td>
<td>36.9</td>
<td>223</td>
<td>0.56</td>
<td>69.1</td>
<td>0.72</td>
<td>0.75</td>
</tr>
<tr>
<td>November 1986</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>November 1985 to</td>
<td>375</td>
<td>(14)</td>
<td>38</td>
<td>0.50</td>
<td>11.8</td>
<td>0.77</td>
<td>1.01</td>
</tr>
<tr>
<td>November 1986</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>February–March 1986</td>
<td>61</td>
<td>(14)</td>
<td>33</td>
<td>0.06</td>
<td>11.7</td>
<td>0.68</td>
<td>0.83</td>
</tr>
<tr>
<td>April–June 1986</td>
<td>175</td>
<td>(14)</td>
<td>41</td>
<td>0.11</td>
<td>16.2</td>
<td>0.71</td>
<td>0.55</td>
</tr>
<tr>
<td>October 1986</td>
<td>38</td>
<td>(14)</td>
<td>36</td>
<td>0.01</td>
<td>11.1</td>
<td>0.74</td>
<td>1.14</td>
</tr>
<tr>
<td>Tracer</td>
<td>25</td>
<td>NA</td>
<td>14</td>
<td>0.00</td>
<td>16.0</td>
<td>0.19</td>
<td>0.96</td>
</tr>
</tbody>
</table>

NA, not applicable.
Fig. 5. Illustration of (a) the mixing cell model and (b) the change in kernel function shape with increasing number of cells (adapted from Himmelblau and Bischoff [1968]).

model with $l$ mixing cells is found by solving (17) for an instantaneous input into cell 1 under initial conditions of no tracer present

$$c_{i}(0) = 0 \quad c_{i-1}(t) = c^{*}(t)[V/Q]\delta(t)$$

where $c^{*}(t)$ is the average concentration of tracer if distributed throughout system. $V$ is the total volume of all cells and is equal to $IV'$, and $\delta(t)$ represents a dirac delta function.

Figure 5b schematically illustrates shapes of kernel functions for systems with different degrees of mixing modeled by a series of mixing cells. As the number of cells in the model increases, the degree of mixing decreases. Time moments of kernels of the mixing cell model also depend on the number of mixing cells [Himmelblau and Bischoff, 1968]

$$\mu_n = \frac{(n + l - 1)(n + l - 2) \cdots (l)}{l^n}$$

and $C_v$ is

$$C_v = (1/l)^{1/2}$$

For the ideally mixed case, $l = 1$, and $C_v$ is 1. For plug flow, $l \to \infty$ and $C_v \to 0$.

Kernel functions and time moments for real aquifers fall within these two extremes. Kernel functions for conduit flow aquifers should have relatively low variance and fit models with a large number of mixing cells. For example, the $C_v$ of the kernel for the tracer test at Maramec Spring is 0.19 and therefore from equation (19), the tracer test results will fit a mixing cell model with approximately 27 cells.

Although the mixing cell model oversimplifies transport in karst aquifers, it illustrates the relationship of the derived kernel functions and time moments to the diffuse and conduit flow classification scheme. Clearly, the variance of the kernels represents the degree of mixing that occurs within the aquifer. The higher the variance of the kernel, the lower the number of mixing cells required to fit the aquifer response, and the greater the degree of mixing in the aquifer. Similarly, the lower the variance, the closer the aquifer approaches the conduit-type end-member in the classification scheme.

**Dual-Porosity Transport Model**

A more physically detailed model might consider the karst aquifer as dual-porosity media composed of regions of relatively mobile and immobile storage. The mobile storage is the conduit network and immobile storage includes fine fractures, pores, and dead-end cavities. Water and associated solutes travel rapidly in karst conduits, but some transfer occurs between the conduits and adjacent fractures and pores. This process may have a similar effect on solute transport through the conduit network as observed in soils with immobile phases. At short distances, tracer tests in soil columns exhibit early arrival times and subsequent tailing in the breakthrough curves. At greater distances, the breakthrough curves have more regular shapes and larger associated dispersion coefficients.

The time moments of the tracer test kernel can be used to compute regional scale, effective transport parameters by, first, assuming a model for convective-dispersive transport, then
comparing the empirical moments to the moments of the impulse response of the model. De Smedt and Wierenga [1984] describe one-dimensional transport in dual-porosity media in the context of flow in unsaturated soils. As an example, we apply their approach here to interpret the kernel function from the Maramec Spring tracer test.

For one-dimensional flow through porous media with mobile and immobile regions, De Smedt and Wierenga [1984] describe transport in the mobile fraction by the expression

$$\frac{\partial C_m}{\partial T} = \frac{1}{P_m} \frac{\partial^2 C_m}{\partial X^2} - \frac{\partial C_m}{\partial X} - \frac{C_m - C_{im}}{P_{im}}$$

and a simple mass transfer relationship for transfer into and out of stagnant zones

$$(1 - \phi) \frac{\partial C_{im}}{\partial T} = \frac{C_m - C_{im}}{P_{im}}$$

where $X = z/L$, $T = vt/L$, $P_m = v_m L/D_m$, $P_{im} = v_{im} L/\theta_m$, and $\phi = \theta_m/\theta$; $\theta_m$ and $\theta_{im}$ are volumetric fractions of mobile and immobile water content, and $C_m$ and $C_{im}$ are solute concentrations in mobile and immobile water. $D_m$ is the hydrodynamic dispersion coefficient in the mobile phase, $v_m$ is the average mobile flow velocity, $\alpha$ is a mass transfer coefficient, and $L$ is the length of the porous media. The distance along the flow path is $z$ and $v$ is average seepage velocity.

Passioura [1971] demonstrated that for flow in cylindrical pore spaces whose length is much larger than their radius, the mass transfer equation is not necessary if the mobile phase dispersion coefficient is redefined to include effects of both mobile and immobile transfer. Solute transport in the mobile phase can then be approximated by

$$\frac{\partial C_m}{\partial T} = \frac{1}{P_{eff}} \frac{\partial^2 C_m}{\partial X^2} - \frac{\partial C_m}{\partial X}$$

where

$$\frac{1}{P_{eff}} = \frac{1}{P_m} + (1 - \phi)^2 P_{im}$$

From the definition of the Peclet numbers, the effective dispersion coefficient $D_{eff}$ is

$$D_{eff} = \alpha D_m + (1 - \phi)^2 v_{eff}^2$$

where $v_{eff}$ is the average effective flow velocity, $v_{eff} = v_m \phi$. De Smedt et al. [1981], Baker [1977], Rao et al. [1980], and others have applied this approach to transport through soils and aggregated porous media with variable pore sizes.

Effective transport properties for the karst aquifer can be computed by comparing the first and second moments of the tracer test kernel function to the moments of the unit impulse response of the dual-porosity model. Following the approach of Aris [1958] and Valocchi [1985], the first and second moments of (23) are found to be

$$\bar{t} = \bar{x}/v_{eff}$$

$$\sigma^2 = 2D_{eff}\bar{t}/v_{eff}$$

where $\bar{x}$ is the average linear travel distance of a solute or tracer.

The $D_{eff}$ and $v_{eff}$ for a given travel distance can be computed directly from the empirical values of $t$ and $C_r$. For example, for an approximate travel distance of 21 km between the tracer input location and Maramec Spring, $v_{eff} = 1.3$ km/day and $D_{eff} = 0.38$ km$^2$/day. The effective dispersivity, defined as

$$\alpha_{eff} = D_{eff}/v_{eff}$$

is 0.29 km. This value is about orders of magnitude smaller than the dimensions of the spring recharge basin, which is approximately 20 km wide and 50 km long. $D_{eff}$ and $v_{eff}$ are regional scale, temporal, and spatial averages that do not represent the discrete features in the karst aquifer. Thus a convective-dispersive model using these regional scale properties should be applied with caution when simulating transport in the karst conduit system at short times or close to the source of contamination. Nevertheless, the effective parameters provide an additional means for comparing transport processes between different locations within a single karst aquifer or among different karst aquifers.

**SUMMARY AND CONCLUSIONS**

A number of observations and conclusions can be made from analysis of the derived kernel functions and a comparison of the functions to alternative models for transport in karst aquifers.

1. A single linear kernel function is sufficient to simulate the storm-derived component of Maramec springflow for both long- and short-term storm sequences. Thus rapid transport to the spring of nonpoint source precipitation is well-approximated as stationary in time.

2. Ca concentrations in the secondary porosity component of Maramec spring flow appear to vary seasonally. This variation may be the result of seasonal mixing of rapidly recharged, relatively dilute water with preexisting water in less mobile storage in fractures, pores, and dead-end cavities.

3. The kernels derived from the spring flow storm responses represent the residence time distribution of rapid groundwater recharge in the conduit network. Because the recharge source is areally distributed, the kernels describe regional scale transport over numerous pathways and a distribution of travel distances. The kernel function derived for a tracer test at Maramec Spring has a larger mean residence time and much smaller variance than the storm response kernels. This implies that the travel distance of the tracer was longer than the mean travel distances of rapid infiltration and that much of the variance in the nonpoint source kernels results from the distribution of flow path lengths to the spring. The skewness values for the kernels are similar, suggesting that the skewness is not sensitive to the distribution of travel path lengths.

4. The time moments of the kernels are a convenient means of describing mean travel times and the degree of mixing that occurs during rapid transport in the aquifer. Kernels for conduit-type aquifers, such as the one studied here, should have a relatively low variance, reflecting the fact that little mixing takes place during transport.

5. Regional scale, effective transport properties can be computed directly from the moments of a tracer test kernel by assuming an effective transport model and comparing the moments of the empirical kernel to the moments of the impulse response of the model. Following this procedure for the Maramec tracer test, we computed an effective velocity between the tracer input point and Maramec Spring of 1.3 km/day and an effective dispersivity of 0.29 km.
6. The principal advantage of the time moment approach is that it describes regional scale transport through a karst aquifer under natural conditions using readily measured spring flow properties. This approach should prove to be an informative method for studying and comparing transport in karst aquifers.

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